
SEASONAL VARIATION OF VOLATILE POLY AROMATIC HYDROCARBONS (PAHS) RELEASED FROM DIFFERENT SOURCES IN SOUTH CAIRO

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ABSTRACT

Understanding the seasonal variations of polycyclic aromatic hydrocarbons (PAHs) concentrations in ambient air in urban/rural or industrial regions is important to the effective control of air pollution in these regions. Based on an approximately a year round dataset (from January to November 2014), a total of 48 atmospheric samples were collected by high volume active air sampler over one year seasonal monitoring in order to assess the variation of PAHs concentrations in different areas through different seasons in south Al Tabbin area. The gaseous and particulate phases of PAHs were extracted and analyzed using GC/MS together. The total air concentrations of 16 USEPA PAHs in the study area ranged from $76.48 \pm 19.44 \mu\text{g}/\text{m}^3$ to $26995.86 \pm 2835.91 \mu\text{g}/\text{m}^3$, with an average concentration of $7085.08 \pm 773.98 \mu\text{g}/\text{m}^3$. Seasonal trends of PAH concentrations were observed with high concentration in winter and low in summer where the average concentration of PAHs in winter were ~1.6 times higher than that in summer. This mainly attributed to meteorological conditions.

Key words: Poly Aromatic Hydrocarbons; Seasonal variation; South Tabbin

INTRODUCTION

Polycyclic aromatic hydrocarbons (PAHs) are a large group of ubiquitous persistent semi-volatile organic compounds (SVOCs). They are comprised of two or more fused benzene rings arranged in various configurations. They are a group of carcinogenic and mutagenic pollutants that are resistant to degradation and can remain in the environment for long periods due to their high degree of conjugation and aromaticity. They have a relatively low solubility in water, but are highly lipophilic. Most of the PAHs with low vapor pressure in the air are adsorbed on particles. When dissolved in water or adsorbed on airborne particulate matter, PAHs can undergo photodecomposition when exposed to ultraviolet light from solar radiation. In the atmosphere, PAHs can react with pollutants such as ozone, nitrogen oxides and sulfur dioxide, yielding diones, nitro- and dinitro-PAHs, and sulfonic acids, respectively.

There is much information on the multi-ringed heavier PAHs but have left the lighter vapor-phase PAH components rather neglected. Although these lighter compounds have weaker carcinogenic/mutagenic properties, they are the most abundant in the urban atmosphere and react with other pollutants to form more toxic derivatives (Park *et al.*, 2002). A PAH classification is done based on their number of aromatic rings, which is used to construct the distribution pattern of the PAHs for each fraction. PAHs can be classified by their numbers of aromatic ring as follows: 2-ring including Nap; 3-ring including ACE, ACY, FLO, PHE and ANT; 4-ring including FLA, PYR, BaA and CHR; 5-ring including BbF, BkF and BaP; 6-ring

including DBahA, IcdP and BghiP. They can also be further classified into lower molecular weight (LMW) containing 2 and 3-ring PAHs, middle molecular weight (MMW) containing 4-ring PAHs and higher molecular weight PAHs (HMW) containing 5 and 6-ring PAHs (Yang *et al.*, 1998, 2002; Chen *et al.*, 2003and 2007).

However, in Egypt, Al Tabbin city that is located in south of Cairo governorate and track neighborhood of Helwan city. Its land area is 25 km². The population in the residential area is about 100,000 according to the population count in 2014. This city is considered as one of the largest industrial cities in Egypt where it has the largest factories for the heavy industries such as “Egyptian Iron and Steel company”, “National Cement company”, “Helwan Cement company”, “Nasr Company for Coke and Basic Chemicals”, “Egyptian company for Metallurgical Industries” and “Helwan Fertilizers Company” as well as many bricks plants. Furthermore, it is bordered by two heavy trafficked highways, Nile Cornish and Al Tabbin Autostrad road.

The main objectives of this study is to measure the concentration levels of ambient PAHs, study the spatial and temporal variations of ambient PAH levels and their possible relationships with meteorological parameters inordert to identify and allocate possible sources of PAHs using diagnostic ratio.

METHODOLOGY AND TECHNIQUES

The applied methodology in this study is based on **US-EPA Method TO-13A** (Compendium Method for determination of PAHs in ambient air). Concisely, the method is based on using a High-Volume Air Sampler for collection of PAHs from ambient air onto the sampling module that consists of particle filter and high volume collection tube containing adsorbent media (i.e. sorbent cartridge).

The study area for this study is located in the south of El Tabbin city with geographical coordinates of $29^{\circ}44'59.92''$ N to $29^{\circ}47'35.12''$ N latitude and $31^{\circ}17'35.45''$ E to $31^{\circ}20'12.13''$ E longitude, and total area of 25 km^2 . The selected sampling area is surrounded by a mixture of urban, industrial, commercial and traffic activities as they are possible sources of PAHs specially the coke production plant. Furthermore, this area represents a large urban industrialized area in El Tabbin city and even in Helwan city where metallurgical, chemical, coal, petrochemical, bricks and cement-producing plants are located. The sampling was conducted in four location within the study area they are namely Tabbin Institute (TIMS) , Residential Area (RA) , Coke Factory (CK) and Arab Abu Said (AAS).

The applied method in this study utilized the High-Volume Air Sampler (Andersen Instruments Inc., 500 Technology Ct., Smyrna, GA) system to acquire sufficient sample for analysis. This system is capable of collecting ambient air through the filter/sorbent cartridge at a flow rate of approximately 225 L/min (i.e. $0.225 \text{ m}^3/\text{min}$) to obtain a total sample volume of greater than 300 m^3 over a 24-hour period. Moreover, this method provides efficient

collection of most PAHs involving two member rings or higher either in a particulate phase or in a gaseous phase through utilization of quartz fiber filter with adsorbent cartridge consists of Poly urethane foam (PUF) and XAD-2 polymer resin as a sorbent media where XAD-2® is intermediately between two layers of polyurethane foam (PUF) in sandwiching configuration in order to minimize breakthrough of highly volatile PAHs. The sampling module inside the high-volume sampler is consists of metal filter holder capable of holding a 102-mm circular particle filter supported by a 16-mesh stainless-steel screen and attaching to a metal cylinder capable of holding a 65-mm outer diameter (O.D.) & 60-mm inner diameter (I.D.) x 125-mm in height borosilicate glass sorbent cartridge containing the adsorbent media (PUF and XAD-2®).

The GC-MS instrument is combined of two parts: The gas chromatograph (GC) and the mass spectrometer (MS), for general specifications for Shimadzu - GCMS-QP5050A, it is a benchtop quadrupole mass spectrometer features an extended mass range to 900 Daltons and optional positive and negative chemical ionization (PCI and NCI). The efficient 150 L/sec turbomolecular pump and rotary pump quickly achieve a high vacuum in less than 5 minutes and this will reflect in minimal downtime. For developing calibration curve for GC/MS analysis, the stock standard solution of 16 PAHs mixed in methylene chloride: methanol (50:50) at the concentration of 2000 µg/ml (Sigma Aldrich Chemical Co. Inc. USA) was diluted first to concentration of 25 ng/µL by taking 125 µL of the stock PAH standard and diluting with hexane in a 10-mL volumetric flask. After that, five concentration levels of PAH standards (i.e., 2.50 ng/µL, 1.25 ng/µL, 0.50

ng/ μ L, 0.25 ng/ μ L, and 0.10 ng/ μ L) were prepared prior GC/MS analysis, each 1 mL aliquot of the five calibration standards as well as 1mL portion of the sample extracts are spiked with 10 μ L at concentration of 50 ng/ μ L (after 1:40 dilution) of the mixture of deuterated PAH internal standard (Isotopically labeled PAH isomers at the concentration of 2000 μ g/mL mixed in methylene chloride, Sigma Aldrich Chemical Co. Inc. USA), to yield a final concentration of 0.5 ng. This internal standard was used to quantify and correct the amounts of specific PAHs found in the samples.

RESULTS AND DISCUSSION

The atmospheric PAHs mass concentrations for different sampling sites and seasons under study are provided in Tables 1.A–D. As can be seen, the total amounts of analyzed PAHs (i.e. Total \sum 16 PAHs) in the area under study varied from $76.48 \pm 19.44 \mu\text{g}/\text{m}^3$ in RA site to $26995.86 \pm 2835.91 \mu\text{g}/\text{m}^3$ in the CK site with a mean concentration of $7085.08 \pm 773.98 \mu\text{g}/\text{m}^3$. As illustrated in Table 1 A-D, the PAH concentration levels demonstrate that the area under study is influenced by regional sources. In the nutshell, the total \sum 16 PAHs concentrations as well as the average concentrations of \sum 16 PAHs over the seasonal sampling period in the different functional zones of the study area can be ordered as follow: CK site (at the border of Coke production plant) ($26995.86 \pm 2835.91 \mu\text{g}/\text{m}^3$ with the average concentration of $6749 \pm 709 \mu\text{g}/\text{m}^3$) > TIMS site (at the prevailing wind directions carried the combustion emissions coming from stationary sources of several industrial complexes) ($901.34 \pm 166.22 \mu\text{g}/\text{m}^3$ with the average concentration

of $225 \pm 42 \text{ } \mu\text{g}/\text{m}^3$) > AAS site (at ≤ 1 Km northwest of the stationary exhausts of the bricks plants) ($366.64 \pm 74.35 \text{ } \mu\text{g}/\text{m}^3$ with the average concentration of $92 \pm 17 \text{ } \mu\text{g}/\text{m}^3$) > RA site (dual impacts from the vicinity to industrial complexes and vehicle emissions) ($76.48 \pm 19.44 \text{ } \mu\text{g}/\text{m}^3$ with the average concentration of $19 \pm 4 \text{ } \mu\text{g}/\text{m}^3$).



Figure 1: Map of the study area where the red spots indicate sampling sites and the yellow lines indicate national ways.

In comparison with others, Wu *et al.* (2014) pointed out that the total PAHs concentrations associated with PM_{2.5} and PM₁₀ in the E'erduosi city in China decreasing as coal-chemical base site > heavy industrial site > residential site with heavy traffic > suburban site surrounded by grassland > background site where PAHs concentrations in the coal-chemical base site are 250 and 31 times of those in the background site for PM_{2.5} and PM₁₀ and respectively.

For the sake of illustration, figure 2 shows the measured total PAH concentrations (i.e. Total \sum_{16} PAHs) over the seasonal sampling period for the different sampling sites, while figure 4.2 shows the measured total PAHs concentrations for each individual compound (\sum_i PAHs) over the seasonal sampling period for each sampling site. Obviously from both figures, the highest mass concentrations of PAHs among the four sampling sites (either for Total \sum_{16} PAHs or \sum_i PAHs) are found in the Coke samples whereas the lowest values are found in the Residential Area samples.

Kim *et al.* (2013) also indicated that the highest values for atmospheric PAHs are generally found when industrial, rather than traffic or residential, contributions are dominant. On the other hand, Zhang and Tao (2009) concluded that biofuel; wildfire, domestic coal combustion, coke production and open straw burning are the most important sources for global atmospheric PAHs, accounting for 56.7%, 12.4%, 11.7%, 7.0% and 2.7%, respectively. However, the combustion derived PAHs (COMPAHs), including FLA, PYR, CHR, BbF, BkF, BaA, BaP, IcdP, BghiP (Bourotte *et al.*, 2005; Kong *et al.*, 2010) can be used to identify the influence of combustion sources on PAHs.

The total concentrations of COMPAHs in this study for the different sampling sites are ranged from $63.24 \pm 17.35 \text{ } \mu\text{g}/\text{m}^3$ to $17546.97 \pm 1848.55 \text{ } \mu\text{g}/\text{m}^3$ accounting for 65 % – 83 % of total PAHs with the highest mass percentages occurred in RA. The calculated COMPAH/ Σ PAH ratios for different sampling sites are 0.65, 0.61, 0.75 and 0.83 for CK, TIMS, AAS and RA, respectively.

Wu *et al.* (2014) demonstrated that the total concentrations of COMPAHs cover 42% – 84% and 75%–82% of PAHs associated with PM_{2.5} and PM₁₀, with the highest mass percentages found in Dongsheng (DS) site of sampling area of E'erduosi city in china. This site represent a residential site at the central urban adjacent to a heavy-traffic road.

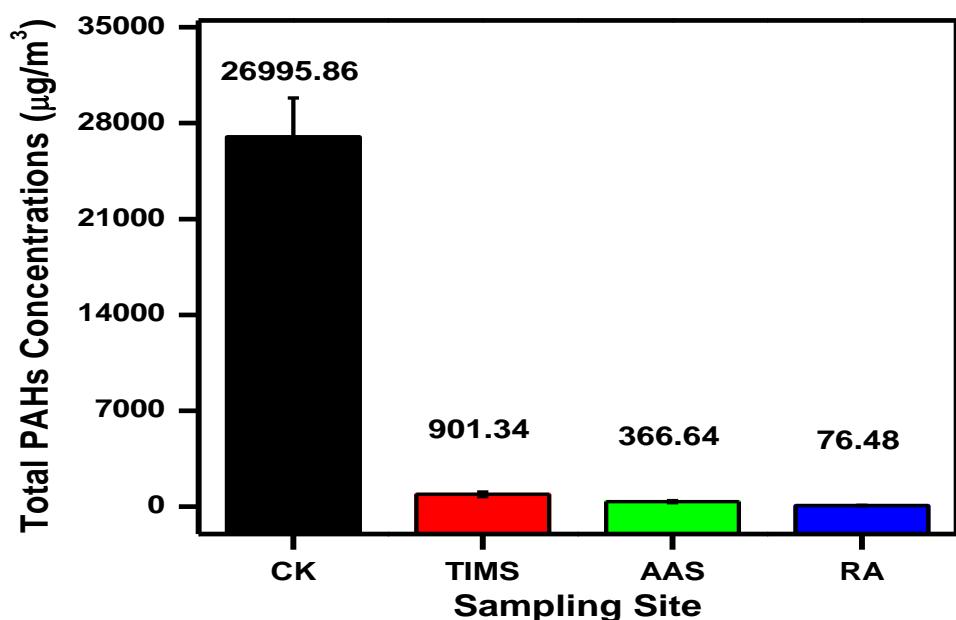


Figure 2: Total PAHs Concentrations over the seasonal sampling period for all sampling sites.

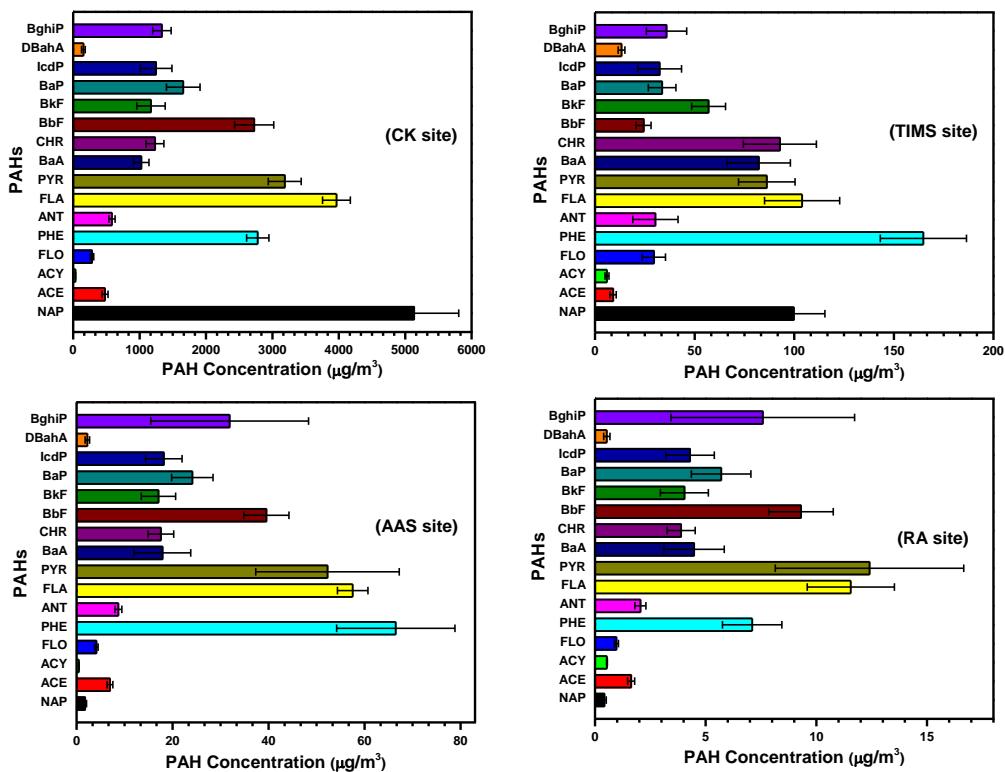


Figure 3: Total of the individual PAH concentrations over the seasonal period for each sampling site.

As an indicator PAHs, BaP was the most carcinogenic PAHs (Johansson and Bavel, 2003). Its content in this study over the seasonal sampling period varied from $5.7 \pm 1.4 \text{ } \mu\text{g}/\text{m}^3$ in RA samples to $1657 \pm 252 \text{ } \mu\text{g}/\text{m}^3$ in CK samples, hold a mean value of $430 \pm 66 \text{ } \mu\text{g}/\text{m}^3$, accounting for ~ 4% in TIMS to ~ 8% in RA of total PAHs concentrations in study area. However, the total mass concentrations of the carcinogenic PAHs (ΣC -PAHs) (including BaA, CHR, BbF, BkF, BaP, IcdP, DBahA) (USEPA, 1999 & 2008) over the

seasonal sampling period in this study are in the range of $32 \pm 7 \text{ }\mu\text{g}/\text{m}^3$ in RA samples to $9213 \pm 1279 \text{ }\mu\text{g}/\text{m}^3$ in CK samples with the average concentrations of $2429 \pm 344 \text{ }\mu\text{g}/\text{m}^3$, accounting for 34–42 % of total PAHs.

As illustrated in Table 1.A-D, since LMW and MMW –PAHS represents the gas phase and HMW –PAHS represent the solid phase the mass concentrations of law macular weight (LMW), medium molecular weight (MMW) and high molecular weight (HMW) -PAHs are in the range of $9294.54 \pm 959.45 - 12.71 \pm 1.95 \text{ }\mu\text{g}/\text{m}^3$, $9409.73 \pm 708.90 - 32.33 \pm 8.22 \text{ }\mu\text{g}/\text{m}^3$, $8291.59 \pm 1167.57 - 31.44 \pm 9.27 \text{ }\mu\text{g}/\text{m}^3$, respectively. Relatively high proportions of MMW-PAHs ($\sim 35\% - \sim 43\%$) and HMW-PAHs ($\sim 31\% - \sim 41\%$) were detected in samples from different sampling sites. Particularly, for CK and TIMS sites, LMW-, MMW-, and HMW-PAHs represent $\sim 34\%$, $\sim 38\%$, $\sim 35\%$ and $\sim 41\%$, $\sim 31\%$, $\sim 22\%$, respectively, of the total PAHs, whereas for AAS and RA sites, the MMW-PAHs were predominantly compounds of the total PAHs (i.e. $\sim 40\%$ and $\sim 43\%$, respectively) followed by HMW-PAHs accounting for $\sim 36\%$ and $\sim 41\%$ then LMW-PAHs (i.e. $\sim 24\%$ and $\sim 17\%$), respectively.

Figure 3 shows the ring distribution of the PAHs over the sampling period for each sampling site. In general, for the area under study, 4-ring PAHs were the dominant form of PAHs measured in samples with the values ranged from 35% to 43%. Then 3-ring and 5-ring PAHs were varying from 15% – 27% and 13% – 25%, respectively. For 2-ring PAHs, the mean percentage for different sampling sites was 8%.

However, results for the ring distribution of PAHs are comparable to others where 2- and 3-ring PAHs are the dominant form of the PAHs measured in coke manufacturing (Khalili *et al.*, 1995; Yang *et al.*, 1998). You (2008) concluded that 4-ring PAHs were most abundant during coal combustion, accounting for about 80%, 70% and 90% at three individual sampling points. Same conclusion was made by Arditoglou *et al.*, (2004) that the PAHs were dominated by 4-ring species (48-62%) followed by 3-ring compounds (38-41%) in fly ashes samples from lignite-fired power plants. 5-ring PAHs exhibited equal or even higher contribution respect to 3- and 4-ring species in coke oven stacks (Manoli *et al.*, 2004). Additionally, Ravindra *et al.*, (2006) indicated that the major source for 3- and 4-ring PAHs is coal combustion, while the major source for HMW PAHs (BaP, BbF, BghiP and Ind) is gasoline vehicles. Therefore, PAHs at the four sites may come mainly from atmospheric transport of the coal combustion and vehicle emissions as both dominated the PAH sources. Regardless, as shown in figure 4, the total PAHs concentrations (i.e. Total \sum 16 PAHs) over the seasonal sampling period for the different sampling sites are well correlated with \sum LMW-PAHs, \sum MMW-PAHs, \sum HMW-PAHs, \sum COMPAHs, \sum C-PAHs over the seasonal sampling period for the different sampling sites with the correlation coefficients higher to 0.99.

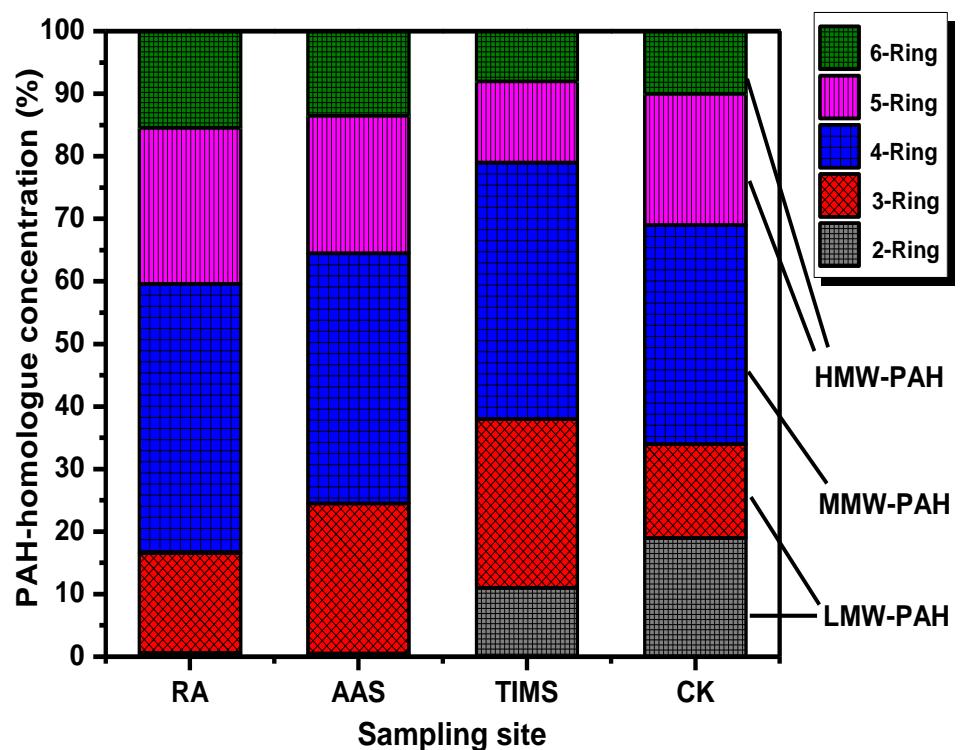


Figure 4. PAH-Homologue concentrations of samples from different sites

Table 1.A. PAHs mass concentrations ($\mu\text{g}/\text{m}^3$) found in the Coke site over the seasonal period

PAHs	Winter	Spring	Summer	Fall	$\sum_i \text{PAH}^b$
NAP	1872.64 \pm 370.06 ^a	1209.88 \pm 95.55	771.68 \pm 42.05	1281.14 \pm 164.08	5135.34 \pm 671.74
ACE	135.82 \pm 12.23	123.98 \pm 7.48	105.07 \pm 20.93	114.53 \pm 4.41	479.40 \pm 45.03
ACY	9.85 \pm 0.93	8.65 \pm 0.16	6.50 \pm 0.72	8.13 \pm 0.36	33.13 \pm 2.17
FLO	77.93 \pm 8.26	64.18 \pm 5.09	58.93 \pm 4.30	82.96 \pm 7.52	284.01 \pm 25.13
PHE	823.67 \pm 14.88	666.35 \pm 29.99	590.27 \pm 48.30	699.07 \pm 74.55	2779.35 \pm 167.72
ANT	151.08 \pm 6.81	132.65 \pm 6.32	149.40 \pm 13.22	150.18 \pm 21.25	583.31 \pm 47.61
FLA	1101.35 \pm 62.46	1017.33 \pm 30.00	853.84 \pm 66.15	993.31 \pm 50.56	3965.83 \pm 209.18
PYR	861.64 \pm 67.52	770.17 \pm 34.90	747.33 \pm 32.33	807.57 \pm 113.34	3186.71 \pm 248.10
BaA	285.67 \pm 36.61	260.31 \pm 22.33	221.24 \pm 29.41	257.36 \pm 29.82	1024.57 \pm 118.12
CHR	348.43 \pm 22.68	311.18 \pm 29.48	276.03 \pm 21.16	296.98 \pm 60.21	1232.62 \pm 133.54
BbF	815.67 \pm 34.15	640.00 \pm 93.58	523.33 \pm 107.34	745.33 \pm 59.97	2724.33 \pm 295.04
BkF	370.57 \pm 57.44	264.00 \pm 44.51	200.33 \pm 46.14	336.67 \pm 66.15	1171.57 \pm 214.25
BaP	540.00 \pm 102.56	413.33 \pm 86.69	312.00 \pm 22.34	391.37 \pm 40.81	1657.00 \pm 252.39
IcdP	401.00 \pm 58.92	324.67 \pm 64.49	230.00 \pm 35.59	292.52 \pm 79.10	1248.33 \pm 238.11
DBahA	43.41 \pm 10.10	39.72 \pm 6.26	30.36 \pm 4.35	40.87 \pm 7.20	154.35 \pm 27.91
BghiP	426.00 \pm 21.79	333.00 \pm 43.97	258.67 \pm 40.07	318.33 \pm 34.03	1336.00 \pm 139.86
$\Sigma_{16} \text{PAHs}^c$	8264.74 \pm 887.42	6579.40 \pm 600.78	5334.98 \pm 534.33	6816.75 \pm 813.37	
LMW-PAHs ^d	3071.00 \pm 413.18	2205.68 \pm 144.59	1681.85 \pm 129.52	2336.01 \pm 272.16	
MMW-PAHs ^e	2597.09 \pm 189.27	2359.00 \pm 116.71	2098.44 \pm 148.97	2355.21 \pm 253.94	
HMW-PAHs ^f	2596.65 \pm 284.97	2014.72 \pm 339.48	1554.69 \pm 255.84	2125.53 \pm 287.27	

^a Mean value (from 3 consecutive sampling weeks) \pm Standard deviation.

^b $\sum_i \text{PAH}$: Summation of the individual PAH concentrations over the seasonal sampling period.

^c $\sum_{16} \text{PAHS}$: Summation of 16 PAHs concentrations for each sampling season.

^d LMW-PAHs: Total Low Molecular PAHs (including NAP, ACE, ACY, FLO, PHE and ANT) for each sampling season.

^e MMW-PAHs: Total Medium Molecular PAHs (including FLA, PYR, BaA and CHR) for each sampling season.

^f HMW-PAHs: Total High Molecular PAHs (including BbF, BkF, BaP, IcdP, DBahA and BghiP) for each sampling season.

ⁱ Total $\sum 16 \text{PAHs}$: Total summation of 16 PAHs concentrations over the seasonal sampling period at the specific sampling site.

Table 1.B. PAHs mass concentrations ($\mu\text{g}/\text{m}^3$) found in Tabbin Institute site over the seasonal period.

PAHs	Winter	Spring	Summer	Fall	ΣiPAH
NAP	33.10 \pm 5.56	22.10 \pm 3.57	17.00 \pm 2.91	27.57 \pm 3.58	99.77 \pm 15.62
ACE	3.17 \pm 0.30	2.01 \pm 0.32	1.39 \pm 0.37	2.51 \pm 0.57	9.07 \pm 1.55
ACY	2.07 \pm 0.40	1.26 \pm 0.26	0.91 \pm 0.07	1.76 \pm 0.26	6.00 \pm 0.98
FLO	9.64 \pm 2.15	6.84 \pm 1.12	5.45 \pm 0.68	7.60 \pm 1.86	29.53 \pm 5.81
PHE	51.32 \pm 6.29	38.91 \pm 3.93	29.80 \pm 4.43	44.80 \pm 7.00	164.82 \pm 21.64
ANT	10.03 \pm 3.47	7.75 \pm 3.12	5.46 \pm 2.25	7.07 \pm 2.53	30.31 \pm 11.37
FLA	33.89 \pm 6.38	25.96 \pm 5.62	16.19 \pm 4.05	27.90 \pm 2.81	103.95 \pm 18.87
PYR	28.26 \pm 3.81	20.01 \pm 2.21	13.58 \pm 4.76	24.41 \pm 3.40	86.25 \pm 14.18
BaA	24.78 \pm 4.26	19.09 \pm 4.47	13.43 \pm 3.25	24.92 \pm 3.90	82.22 \pm 15.88
CHR	27.61 \pm 4.44	22.27 \pm 4.07	16.44 \pm 5.69	26.47 \pm 4.16	92.80 \pm 18.36
BbF	7.66 \pm 0.90	5.87 \pm 0.89	4.13 \pm 0.66	6.73 \pm 1.29	24.38 \pm 3.73
BkF	18.86 \pm 1.81	14.50 \pm 2.39	9.04 \pm 1.84	14.60 \pm 2.45	57.00 \pm 8.49
BaP	11.35 \pm 1.64	8.59 \pm 1.85	5.49 \pm 1.57	8.22 \pm 1.88	33.65 \pm 6.93
IcdP	10.95 \pm 3.07	7.96 \pm 3.18	5.49 \pm 2.35	8.02 \pm 2.38	32.42 \pm 10.98
DBahA	4.52 \pm 0.63	3.44 \pm 0.33	2.14 \pm 0.25	3.18 \pm 0.49	13.28 \pm 1.69
BghiP	12.16 \pm 3.14	9.33 \pm 2.75	5.83 \pm 1.91	8.57 \pm 2.32	35.89 \pm 10.12
$\Sigma 16 \text{ PAHs}$	289.36 \pm 48.24	215.90 \pm 40.06	151.75 \pm 37.04	244.32 \pm 40.87	
LMW-PAHs	109.32 \pm 18.17	78.87 \pm 12.31	60.00 \pm 10.70	91.30 \pm 15.80	
MMW-PAHs	114.54 \pm 18.89	87.34 \pm 16.39	59.64 \pm 17.75	103.70 \pm 14.27	
HMW-PAHs	65.50 \pm 11.19	49.70 \pm 11.37	32.11 \pm 8.58	49.32 \pm 10.80	

Table 1.C. PAHs mass concentrations ($\mu\text{g}/\text{m}^3$) found in Arab Abu Said site over the seasonal period.

PAHs	Winter	Spring	Summer	Fall	$\Sigma i \text{ PAH}$
NAP	0.65 \pm 0.14	0.42 \pm 0.05	0.27 \pm 0.02	0.44 \pm 0.07	1.78 \pm 0.28
ACE	1.97 \pm 0.12	1.80 \pm 0.16	1.52 \pm 0.27	1.66 \pm 0.04	6.95 \pm 0.60
ACY	0.14 \pm 0.02	0.13 \pm 0.01	0.09 \pm 0.01	0.12 \pm 0.01	0.48 \pm 0.04
FLO	1.13 \pm 0.09	0.93 \pm 0.07	0.86 \pm 0.08	1.20 \pm 0.12	4.12 \pm 0.36
PHE	18.29 \pm 1.78	13.95 \pm 4.48	14.81 \pm 3.73	19.43 \pm 2.33	66.48 \pm 12.32
ANT	2.25 \pm 0.10	1.98 \pm 0.09	2.23 \pm 0.20	2.24 \pm 0.32	8.71 \pm 0.73
FLA	15.99 \pm 1.34	14.74 \pm 0.15	12.37 \pm 0.69	14.41 \pm 1.0	57.51 \pm 3.12
PYR	14.85 \pm 5.36	13.18 \pm 3.44	11.50 \pm 3.70	12.74 \pm 2.43	52.27 \pm 14.00
BaA	5.15 \pm 2.26	4.45 \pm 1.22	3.88 \pm 1.67	4.40 \pm 0.77	17.88 \pm 5.34
CHR	5.06 \pm 0.45	4.18 \pm 1.00	4.00 \pm 0.34	4.31 \pm 0.89	17.54 \pm 2.65
BbF	11.84 \pm 0.84	9.31 \pm 1.63	7.57 \pm 1.39	10.80 \pm 0.58	39.52 \pm 4.73
BkF	5.40 \pm 1.00	3.84 \pm 0.76	2.92 \pm 0.75	4.90 \pm 1.07	17.04 \pm 3.51
BaP	7.86 \pm 1.69	6.02 \pm 1.44	4.53 \pm 0.43	5.69 \pm 0.75	24.09 \pm 4.31
IcdP	5.83 \pm 1.02	4.73 \pm 1.06	3.33 \pm 0.56	4.27 \pm 1.27	18.15 \pm 3.81
DBahA	0.63 \pm 0.16	0.58 \pm 0.11	0.44 \pm 0.07	0.59 \pm 0.12	2.25 \pm 0.46
BghiP	10.18 \pm 5.70	8.17 \pm 5.35	6.25 \pm 2.20	7.27 \pm 3.19	31.88 \pm 14.44
$\Sigma 16 \text{ PAHs}$	107.21 \pm 22.07	88.39 \pm 21.01	76.56 \pm 16.02	94.49 \pm 15.26	
LMW-PAHs	24.43 \pm 2.25	19.20 \pm 4.86	19.78 \pm 4.30	25.10 \pm 2.90	
MMW-PAHs	41.04 \pm 9.41	36.55 \pm 5.81	31.75 \pm 6.40	35.86 \pm 5.08	
HMW-PAHs	41.73 \pm 10.40	32.64 \pm 10.34	25.03 \pm 5.31	33.53 \pm 7.28	

Table 1.D. PAHs mass concentrations ($\mu\text{g}/\text{m}^3$) found in the Residential Area site over the seasonal period.

PAHs	Winter	Spring	Summer	Fall	$\sum_i \text{PAH}$
NAP	0.15 ± 0.04	0.10 ± 0.02	0.06 ± 0.01	0.10 ± 0.02	0.42 ± 0.08
ACE	0.46 ± 0.01	0.42 ± 0.06	0.36 ± 0.05	0.39 ± 0.03	1.63 ± 0.16
ACY	0.34 ± 0.03	0.05 ± 0.01	0.03 ± 0.01	0.12 ± 0.01	0.54 ± 0.02
FLO	0.26 ± 0.02	0.22 ± 0.02	0.20 ± 0.03	0.28 ± 0.04	0.97 ± 0.07
PHE	1.98 ± 0.29	1.65 ± 0.33	1.76 ± 0.21	1.71 ± 0.50	7.10 ± 1.34
ANT	0.53 ± 0.03	0.47 ± 0.05	0.53 ± 0.08	0.53 ± 0.09	2.05 ± 0.25
FLA	3.10 ± 0.76	2.96 ± 0.30	2.57 ± 0.54	2.92 ± 0.32	11.56 ± 1.97
PYR	3.53 ± 1.40	3.13 ± 0.50	2.73 ± 1.02	3.01 ± 0.75	12.40 ± 4.26
Baa	1.40 ± 0.32	1.11 ± 0.32	0.93 ± 0.43	1.03 ± 0.19	4.47 ± 1.36
CHR	1.19 ± 0.16	0.98 ± 0.24	0.85 ± 0.05	0.87 ± 0.15	3.89 ± 0.63
BbF	2.79 ± 0.37	2.20 ± 0.52	1.77 ± 0.28	2.54 ± 0.29	9.30 ± 1.46
BkF	1.28 ± 0.23	0.91 ± 0.23	0.69 ± 0.12	1.16 ± 0.32	4.04 ± 1.08
Bap	1.86 ± 0.50	1.43 ± 0.34	1.07 ± 0.17	1.34 ± 0.25	5.70 ± 1.35
IcdP	1.38 ± 0.32	1.12 ± 0.31	0.78 ± 0.10	1.01 ± 0.36	4.29 ± 1.10
DBahA	0.15 ± 0.04	0.14 ± 0.03	0.10 ± 0.02	0.14 ± 0.04	0.53 ± 0.14
BghiP	2.45 ± 1.51	1.92 ± 1.24	1.48 ± 0.55	1.74 ± 0.85	7.58 ± 4.15
$\sum_{16} \text{PAHs}$	22.85 ± 6.25	18.81 ± 5.13	15.90 ± 3.84	18.82 ± 4.22	
LMW-PAHs	3.73 ± 0.39	2.91 ± 0.49	2.94 ± 0.38	3.13 ± 0.69	
MMW-PAHs	9.22 ± 2.81	8.19 ± 1.88	7.07 ± 2.11	7.84 ± 1.42	
HMW-PAHs	9.90 ± 3.05	7.71 ± 2.76	5.90 ± 1.34	7.94 ± 2.11	

Indubitably, the environmental locations of the sampling sites are considered as an important contributor to the concentrations of atmospheric PAHs. Because the sources of PAHs are almost related to combustion, the specific local sources may lead to the differences of PAH concentrations for different sites in a city or area under study. Therefore, the comparison studies of atmospheric PAHs in different functional zones are necessary for establishing effective controlling measures in an urban area. Furthermore, owing to difference of population, traffic density and industry distribution, the PAH contribution from anthropogenic inputs varied in different function zones (Lodovici *et al.*, 2003; Kong *et al.*, 2010; Kim *et al.*, 2013).

Accordingly, the sampling strategy in this study was applied to different functional zones (i.e. CK, TIMS, AAS and RA) in order to determine the total PAHs concentrations in the south of El Tabbin city with 25 km².

CONCLUSION

In this study, 16 PAHs in particulate and gaseous phases (LMW and MMW –PAHS represents the gas phase) (HMW –PAHS represent the solid phase) were identified and quantified in 48 atmospheric air samples from 4 sampling sites in four seasons from January to November 2014 at the south of El Tabbin city in order to determine the concentration levels, seasonal variations and their potential sources. The results showed the mass concentration of PAHs with the highest concentrations at the industrial areas (CK, TIMS and AAS) following by residential area (RA). It was found that PAHs concentrations were high in winter and autumn at the four sampling sites where it take the order of winter > autumn > spring > summer. This

behavior could be attributed to the meteorological factors such as regional climatic conditions, lower atmospheric mixing height, decreased sunlight intensity as well as frequent temperature inversion that intensify the PAH pollution in winter. Atmospheric outflow was estimated based on the concentration of PAHs and wind velocity where the elevated transport fluxes were found during the spring and winter seasons and southeastward transport is dominated.

Among the 16 PAHs, FLA, PYR, CHR, BbF, BkF, BaA, BaP, IcdP, BghiP (represent combustion derived PAHs) were most abundant, which reflects the influence of combustion processes (either coal or oil) and vehicle emission. Predominantly coal combustion, vehicle emission (including diesel and gasoline fuel exhaust) and industrial processes are the primary sources for atmospheric PAHs in this heavy-industrialized city. BaP concentration in the study area is extremely high where the average BaP concentration over seasonal sampling period for the study area, reflecting a serious hidden danger to health.

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التأثيرات الموسمية للمركبات العضوية المتطايرة عديدة المطاقم الناتجة من المصادر المختلفة

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المستخلص

من المهم فهم التغيرات الموسمية المركبات العضوية عديدة الحلقات في المناطق الحضرية والصناعية والسكنية من أجل التحكم الفعال في ثلوث الهواء في تلك المناطق وبناء على بيانات ودراسات استمرت لعام كامل (من يناير إلى نوفمبر ٢٠١٤) تم سحب عدد ٤٨ عينة باستخدام جهاز السحب على القدرة تم سحب مجموعة من العينات بصورة مكثفة في تلك الفترة في منطقة جنوب التبين وذلك من أجل تقييم التغيرات الموسمية لتلك المركبات في تلك المنطقة حيث تم سحب عينات من المركبات العضوية عديدة الحلقات في حالتها الصلبة والغازية حيث تم استخلاص المركبات العضوية عديدة الحلقات وتم تحليلها باستخدام جهاز الفصل الكروماتوجرافى الغازى والذي يحتوى على جهاز تحليل طيف الكتلة. تراوحت تركيزات المركبات العضوية عديدة الحلقات والمصنفة طبقاً لتعليمات هيئة حماية البيئة الأمريكية بين $19,44 \pm 76,48$ ميكروجرام لكل متر مكعب إلى 2699 $5,86 \pm 2835,91$ ميكروجرام لكل متر مكعب وبمتوسط تركيز $773,98 \pm 7080,08$ ميكروجرام لكل متر مكعب. اختلفت تركيزات المركبات العضوية عديدة الحلقات باختلاف موقع القياسات.

تم تحديد التغيرات الموسمية للمركبات العضوية عديدة الحلقات حيث جاءت التركيزات عالية في موسم الشتاء والتركيزات قليلة في موسم الصيف حيث جاءت نتائج فصل الشتاء حوالي ١,٦ ضعف نتائج فصل الصيف والذي من الممكن تقسيمه باستخدام العوامل المناخية.